

REPORT NO. 2503

TECHNICAL REPORT OF MACROINVERTEBRATE COMMUNITY INDEX PREDICTIONS FOR THE WELLINGTON REGION

TECHNICAL REPORT OF MACROINVERTEBRATE COMMUNITY INDEX PREDICTIONS FOR THE WELLINGTON REGION

JOANNE CLAPCOTT, ERIC GOODWIN

Prepared for Greater Wellington Regional Council

CAWTHRON INSTITUTE 98 Halifax Street East, Nelson 7010 | Private Bag 2, Nelson 7042 | New Zealand Ph. +64 3 548 2319 | Fax. +64 3 546 9464 www.cawthron.org.nz

REVIEWED BY: Roger Young

Nor $5-1$

APPROVED FOR RELEASE BY: Natasha Berkett

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1. INTRODUCTION

1.1. Background

Greater Wellington Regional Council (GWRC) is currently reviewing its five regional plans and is developing a single new integrated plan for the Wellington region. A discussion document (Regional Plan Working Document for Discussion or WDFD) was released to key stakeholders in August 2013 and a Draft Regional Plan is expected to be released around September 2014.

Macroinvertebrates have been identified as a key indicator of river and stream ecosystem health in Schedule H of the WDFD. However, due to a lack of robust information on variation in reference condition (natural state) of macroinvertebrate metrics in the Wellington region, numeric thresholds¹ were not included. The purpose of this project is to develop a predictive model of contemporary macroinvertebrate metric scores specific to the Wellington region that could then be used to inform numeric thresholds for the Macroinvertebrate Community Index (MCI) to be included in the Draft Regional Plan. A model developed specifically for the Wellington region is likely to provide greater accuracy than the national model (Clapcott *et al.* 2013) as additional data not used in the development of the national model will be used. Predictions from the regional model can then be used in combination with existing reference data to identify reference thresholds for selected classes in the region.

1.2. Project scope

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The purpose of this project is to:

- Develop a model of contemporary MCI scores specific to the Wellington region based on MCI data from 290 sites across the region.
- Compare the accuracy of predictions from the regional model with that from the national model.
- Use the outputs from the regional model along with existing reference data as the basis for MCI thresholds for classes in the Wellington region in four bands — 'excellent', 'good', 'fair' and 'poor'.

The methods for identification of MCI thresholds should be, as far as possible, consistent with any likely future guidance from the National Objectives Framework.

 1 In this instance, 'thresholds' refers to numeric guideline values that can be used to differentiate between specific states of ecological quality, such as reference or 'excellent', 'good', 'fair' or 'poor' quality.

2. METHODS

2.1. Source data

Data was provided by Greater Wellington Regional Council compiled from a range of projects conducted over the last 15 years. Data included MCI scores for 290 sites in the Wellington region; scores were either an average of MCI scores from three replicates or a score from a single replicate at sites sampled on a single occasion, or 5 year medians from routine monitoring sites (Table 1). The 290 sites were relatively representative of the environmental variation observed the Greater Wellington region (Table 2). Sites sampled and model outputs are summarised by two levels of river classification. The first level is the GWRC $FENZ²$ classification that is based on the FENZ classification and documented in Warr (2009). The second is the Regional Plan River Class (RP River Class) which is a further condensed version of the GWRC FENZ classification (Table 2).

Table 1. Summary of Macroinvertebrate Community Index (MCI) sample data analysed in this project.

*Project details:

- A. GWRC Rivers State of the Environment (RSoE) monitoring
- B. GWRC Historic RSoE monitoring
- C. GWRC Mangatarere Stream investigation
- D. GWRC Riparian restoration monitoring
- E. Kingett Mitchell REC verification study
- F. Kingett Mitchell Wellington City urban stream study
- G. GWRC Wellington city urban stream samples
- H. Kingett Mitchell Wairarapa urban streams study
- I. Massey University samples
- J. GWRC additional sampling
- K. GWRC Pahaoa River investigation

 2 Freshwater Ecosystems of New Zealand (FENZ) geo-database

Table 2. Distribution of sample data sites in relation to Greater Wellington Regional Council (GWRC) river classification.

*Description of RP River Classes:

1. Steep gradient, hard sedimentary

2. Moderate gradient and coastal, hard sedimentary

3. Moderate gradient, soft sedimentary

4. Low gradient, large, draining ranges

5. Low gradient, large, draining plains and eastern Wairarapa

6. Low gradient, small

2.1. Predictive model development

We investigated two model approaches recently used to predict contemporary and reference values for stream metrics: boosted regression tree (Clapcott *et al.* 2013) and linear regression (Unwin *et al.* 2010).

2.1.1. Boosted regression tree model

Boosted regression tree (BRT) models were developed using the training data set (n = 290) and 18 environmental predictor variables from the LCDB3³ and $FENZ²$ database (Table 3). Boosted regression tree model-fitting parameters, such as the

 3 New Zealand's Landcover Database v3.0 (LCDB3)

tree complexity and learning rate, were tuned manually with the aim of improving the percentage of deviance explained by the model. A number of alternative models were investigated, such as modelling transformed MCI, including or excluding RP River Class category, and automated simplification of the model by the removal of variables with low contribution. Additionally, we used a leave-one-out cross validation approach which allowed us to independently test the predictive performance of the models and assess model consistency and bias.

The best-performing model was one that modelled logit-transformed MCI, where:

$$
logit(MCI) = \log(\frac{MCI/200}{1 - MCI/200})
$$

All 18 predictor variables were retained in this best performing model which explained 65.9% of the deviance in MCI data and had a cross validation coefficient of 0.818 (se = 0.021). The proportion of native vegetation in the catchment explained over half of the deviance in MCI data and had a strong positive relationship with MCI (Figure 1).

Figure 1. Shape of the relationships between Macroinvertebrate Community Index (MCI) and individual environmental predictors in order of model importance (12 most important variables shown) from a boosted regression tree (BRT) model. Note the rug plots on the x-axis show the distribution of training data and the y-axis scale shows the marginal contribution of each predictor in logit-transformed units to mean MCI.

We compared predictions for the 290 site training dataset from the regional BRT model and the national model from Clapcott *et al.* (2013). Firstly, the regional model shows excellent predictive accuracy (Nash-Sutcliffe efficiency [NSE] = 0.82) and effectively no bias, and the root mean square deviation (RMSD) was 11.05 suggesting a standard deviation of 21.6 MCI units for the model (Figure 2) . In comparison,

predictions from the national model versus the observed values from 290 sites had good predictive accuracy (NSE = 0.69), low bias (0.89) and the RMSD was 14.41 suggesting a standard deviation of 28.2 MCI units.

Figure 2. Correlations between observed and predicted values from the boosted regression tree (BRT) a) regional model and b) national model for contemporary Macroinvertebrate Community Index (MCI). $N = 290$. The dashed line is the 1:1 line and the blue line is the line of best fit. Model performance statistics are explained in text.

The relationships between predicted and observed MCI values from the BRT model using a leave-one-out approach showed good model validation (Figure 3) in terms of:

- the Nash-Sutcliffe efficiency (NSE = 0.67) statistic which indicates how well the plot of observed versus predicted fits the 1:1 line, where values greater than 0 are satisfactory but values greater than 0.5 indicate good model performance,
- root mean squared deviation (RMSD = 14.73) is an estimate of model uncertainty (overall departure between observed and predicted values), where smaller values indicate lower uncertainty than large values,
- bias (Bias = -0.07) which measures the average tendency of the predicted values to be larger or smaller than the observed, where positive values indicate model underestimation and negative values indicate overestimation bias.

These model diagnostics can be interpreted as indicative of a very good predictive model (within the training data range), with $95th$ percent confidence intervals of ≤ 29 MCI units, and effectively no bias (< 0.1 MCI unit). The difference between the potential predictive performance (Coefficient of variation [CV] correlation of 0.82) and

the validation model performance (NSE = 0.67) suggests that the model could be improved by increasing sample N within the environmental range of the training data.

Figure 3. Correlations between observed and predicted values from the boosted regression tree (BRT) leave-one-out validation model for contemporary Macroinvertebrate Community Index (MCI). The dashed line is the 1:1 line and the blue line is the line of best fit. Model performance statistics are explained in text.

2.1.2. Linear model

We explored linear models to provide a comparison to the BRT approach and to test an alternative approach to predicting reference state (discussed below). Linear models were developed using the training dataset (n = 290) and a selection of the 18 environmental predictor variables from the LCDB3 and FENZ database (Table 3). Initially we trialled general additive models (GAM) to allow for up to three-way interactions between smoother functions of all predictors. Inspection of the resultant GAM model suggested that only low-order smoothing functions were being fitted, meaning there was little justification for moving away from a standard linear

regression. Therefore we developed a linear model using manual predictor selection informed by variance inflation factor (VIF) and the Akaike information criterion (AIC) value. We used backward selection ('step' function in core R package 'stats'; R Core Team, 2013) to inform the most parsimonious model. The 'step' function sequentially removes the least significant predictor variable until the model with the lowest AIC value is reached. We also explicitly investigated the potential for interactions of interest (e.g. between *NativeVeg* and environmental descriptors). The 'best' model contained seven predictor variables (Table 3) as well as an interaction between *Urban* and *NativeVeg* and had an adjusted R^2 -value of 0.70 (p < 0.001).

The relationships between predicted and observed MCI values from the linear model using a leave-one-out approach showed good model validation (Figure 4) in terms of:

- \bullet the Nash-Sutcliffe efficiency (NSE = 0.70),
- root mean squared deviation (RMSD = 14.2), and
- \bullet bias (Bias = -0.22).

These model diagnostics can be interpreted as indicative of a very good predictive model, with 95th percent confidence intervals of < 28 MCI units, and effectively no bias (<0.3 MCI units).

Figure 4. Correlations between observed and predicted values from the linear leave-one-out validation model for contemporary Macroinvertebrate Community Index (MCI). The dashed line is the 1:1 line and the blue line is the line of best fit. Model performance statistics are explained in text.

2.1.3. Predicting reference state

To predict expected reference MCI values for the Wellington reaches, the land-use variables were reset to values representative of no human pressure (0 for *DailyAll*, *PastoralHeavy*, *PastoralLight*, *Urban* and *ExoticVeg* and 100 for *NativeVeg*). The BRT and linear model were then used to predict MCI values in the absence of land-use pressure. Boosted regression tree models do not 'extrapolate' beyond the fitted range of predictor variables. So if the resetting of land use creates sites with a combination of characteristics beyond that which was in the training set, BRT models can only suggest the MCI value for the most similar site. By contrast, linear regressions can extend the trend observed along gradients of predictive conditions, and propose an MCI value outside the range experienced in the training set. A comparison of BRT and linear model reference predictions illustrates how BRT predictions are truncated to a value observed in the training data set (Figure 5).

Reference MCI predictions

Figure 5. Correlation of predicted reference Macroinvertebrate Community Index (MCI) values from the boosted regression tree (BRT) and linear models.

A leave-one-out cross validation illustrates the weakness of both model approaches for accurately predicting reference⁴ MCI values on a site-by-site basis (Figure 6). The NSE values reflect the limited validation data set, whereas the RMSD is consistent with that observed for contemporary models. Improved model performance is likely to be gained by increasing N, especially in the reference range.

 4 'Reference' sites having no human pressure (0 for *DailyAll*, *PastoralHeavy*, *PastoralLight*, *Urban* and *ExoticVeg* and 100 for *NativeVeg*)

Figure 6. Correlations between observed and predicted values from the a) boosted regression tree (BRT) and b) linear leave-one-out validation model for reference Macroinvertebrate Community Index (MCI). The dashed line is the 1:1 line and the blue line is the line of best fit. Model performance statistics are explained in text.

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3.1. Output

The output from the BRT and linear models was used to predict MCI values for all stream segments in the Wellington region and to calculate summaries by two levels of stream classification: GWRC FENZ Class and RP River Class (Appendix 1). Boxplots for RP River Class summaries are shown in Figure 7 and Figure 8, and for GWRC FENZ Class see Appendix 2. In terms of contemporary predictions the modelled output distributions mirror the observed, with significantly different ranges of values within the six RP River Classes (Figure 7). For reference predictions, there is a distinct tendency for higher values to occur at streams with lower temperature and higher slopes. This is particularly evident for linear model predictions because *SEGJANAIRT* and *SEGSLOPESQ* were the primary variables driving this model in the absence of land-use pressure (Figure 8). The BRT model incorporated more environmental variation leading to generally lower reference predictions at sites characterised as soft sedimentary (Figure 8).

Figure 7. Boxplots showing measured Macroinvertebrate Community Index (MCI) and contemporary MCI from linear and boosted regression tree (BRT) model predictions grouped by RP River Class.

Figure 8. Boxplots showing measured Macroinvertebrate Community Index (MCI) and reference MCI from linear and boosted regression tree (BRT) model predictions grouped by RP River Class.

3.2. Management bands

Three alternative ways to determine management bands for MCI are discussed below.

3.2.1. Deviation from reference approach

Bands are determined by applying a set deviance from reference approach. For example, if the expected reference MCI value for a site was 130, then bands could be within 20 MCI units of the reference value or within 20% of the reference value (either approach is valid given the normal distribution of MCI data); within 20 MCI units = 'Excellent' (MCI >130), 'Good' (110-129), 'Fair' (90-109) or 'Poor' (<90), or within 20% 'Excellent' (MCI >130), 'Good' (104-129), 'Fair' (78-104) or 'Poor' (<78). However, this approach could lead to very low bottom lines where there are low-scoring reference sites and this may not give sufficient environmental protection.

3.2.2. Ratio of observed to expected (reference)

The ratio of observed to expected (MCI_{OE}) requires a robust estimate of reference on a site by site basis. The current predictive models, whilst very good in terms of explaining deviance in MCI have an RMSD (predictive error) of approximately 15 MCI units. Which means 80% confidence is only achieved when bands are 28 MCI units wide. MCI_{O/F} could be calculated using an average E or reference value for a given stream class. However, using the average reference conditions for stream classes still means there could be large variation between that statistic and the actual reference condition of a particular site. Hence, at the site level, using average reference predictions for different stream classes could have the effect of allowing more degradation (if the average reference value is lower than the actual reference value) or less degradation (if the average is greater than the actual) when compared to using an accurate site-specific reference value. In reality, an average / quantile reference value for a given stream class is the best available option in the absence of sitespecific reference values.

3.2.3. Adoption of existing quality classes

Four quality classes were assigned by Stark and Maxted (2007b) to the MCI to denote 'Excellent' (MCI ≥ 120), 'Good' (100–119), 'Fair' (80–99) or 'Poor' (< 80) conditions indicative of different levels of pollution. The 'Poor' threshold of 80 was calculated as halfway between the 'Excellent' threshold of 120 and the theoretical MCI minimum of 40. Stark and Maxted (2007a) tested these boundaries for soft-bottomed streams and found they corresponded well for the MCI_{sb} which provided greater discrimination of effects for urban and moderate-high intensity rural land uses. Wright-Stow and Winterbourn (2003) proposed that 'fuzzy' rather than fixed boundaries would provide more certainty in assigning quality classes, such that 'Excellent' would become > 125, 'Good' 105–115, 'Fair' 85–95 and 'Poor' < 75, and intermediate values would have

intermediate class assignments (*e.g.* 96–114 would be 'Good-fair'). These fuzzy classes were originally proposed by Stark (1985) and were later replaced by the fixed boundaries, although the concept of fuzzy boundaries was supported by Stark and Maxted (2007b).

The assigning of management band thresholds has been discussed as part of the proposed inclusion of MCI as a NOF variable (NZFSS submission 4 February 2014 http://freshwater.science.org.nz/pdf/NZFSS_amendments_to_the_NPS_FM.pdf). The proposed national bands are the default quality classes assigned by Stark and Maxted (2007b) but an important aspect is noted: "…the statistic used (*e.g*. summer value, monthly mean, 3-year rolling mean) will need to be determined." Analysis of regional data provides improved resolution for the definition of band thresholds for GWRC, and the MCI statistic could be defined by analysis of reference site variability. Analysis of year-to-year variability in MCI values at reference sites in the Waikato region (Clapcott unpublished data) suggests a 3-year rolling mean would be appropriate. This reduces the likelihood that a site will be assigned to the wrong band due to sampling error or natural deviation in MCI values due to climatic influences.

At the regional level, the definition of an 'Excellent' threshold is best defined firstly by the measured MCI value at reference sites by stream class. The number of measured sites required per class can be informed by the variability observed in environmental descriptors. For example, predictive models suggest greater environmental variability in RP River Classes 1, 2 and 3 compared to lower variability in RP River Classes 4 and 5. Data from at least three sites per class could be used in a power analysis to determine the minimum number of sites required to characterise reference state.

Secondly, an 'Excellent' threshold could be defined by the modelled contemporary MCI value for sites that meet a set of land-use filters and thirdly, by the modelled reference value by stream class. In this project we used a set of land-use filters that allowed for a reference data set greater than 10% of the total data set, *i.e.* 40 out of 290 sites. As such, filters were conservative allowing up to 15% non-native vegetation in the catchment and 5% light pasture, but no heavy pasture and no urban development or surface water allocation. Note: using a land-use restriction of 90% native forest reduced the reference data set to 24 sites.

The average or the $25th$ percentile value could be used to inform a reference threshold. The latter would mean accepting that 25% of all reference sites would fall below the 'Excellent' threshold (compared to 50% for the mean value). The definition of remaining thresholds could be achieved by the equal distribution of data from the 'Excellent' threshold down to a lower limit determined by either the $5th$ percentile of measured or modelled data. The latter would mean accepting that measured data may not represent the true range of conditions present in any given class.

3.2.4. RP River Class 1

Summary model statistics for RP River Class 1 (Steep gradient, inland, hard sedimentary) are provided in Table 4. Data suggest an 'Excellent' threshold of 130 to 140 and a 'Fair' threshold of 110. Accepting an 'Excellent' threshold at 140 would suggest a 'Good' threshold halfway at 125 — thresholds at 140, 125, 110.

Table 4. Summary Macroinvertebrate Community Index (MCI) statistics for RP River Class 1. *from the contemporary model for sites with native vegetation > 85%, light pasture < 5%, heavy pasture = 0% , urban = 0% , surface water allocation = 0.

3.2.5. RP River Class 2

Summary model statistics for RP River Class 2 (Moderate gradient and coastal, hard sedimentary) are provided in Table 5. Data suggest an 'Excellent' threshold of 130 and a 'Fair' threshold between 70 and 90; 80 would be in line with the proposed national C/D threshold for hard-bottomed streams (NZFSS submission 4 February 2014). Accepting an 'Excellent' threshold at 130 would suggest a 'Good' threshold approximately halfway at 105 — thresholds at 130, 105, and 80. Note the wider range (minimum to maximum values) in observed and predicted contemporary values for RP River Class 2 compared to RP River Class1. This is likely to reflect the greater range of both environmental and land use gradients in RP River Class 2.

Table 5. Summary Macroinvertebrate Community Index (MCI) statistics for RP River Class 2. *from the contemporary model for sites with native vegetation > 85%, light pasture < 5%, heavy pasture = 0% , urban = 0% , surface water allocation = 0

3.2.6. RP River Class 3

Summary model statistics for RP River Class 3 (Moderate gradient, soft sedimentary) are provided in Table 6. Data suggest an 'Excellent' threshold of 130 and a 'Fair' threshold of 80. Accepting an 'Excellent' threshold at 130 would suggest a 'Good' threshold approximately halfway at 115 – thresholds at 130, 105, and 80.

Table 6. Summary Macroinvertebrate Community Index (MCI) statistics for RP River Class 3. *from the contemporary model for sites with native vegetation > 85%, light pasture < 5%, heavy pasture = 0% , urban = 0% , surface water allocation = 0

3.2.7. RP River Class 4

Summary model statistics for RP River Class 4 (Low gradient, large, draining ranges) are provided in Table 7. Data suggest an 'Excellent' threshold of 120 to 130 and a 'Fair' threshold of 90. Accepting an 'Excellent' threshold at 130 would suggest a 'Good' threshold approximately halfway at 110 — thresholds at 130, 110, and 90. The utility of model predictions is illustrated in this and following RP River Classes where there are no measured data from reference sites.

Table 7. Summary Macroinvertebrate Community Index (MCI) statistics for RP River Class 4. *from the contemporary model for sites with native vegetation > 85%, light pasture < 5%, heavy pasture = 0% , urban = 0% , surface water allocation = 0

3.2.8. RP River Class 5

Summary model statistics for RP River Class 5 (Low gradient, large, draining plains and eastern Wairarapa) are provided in Table 8. Data suggest an 'Excellent' threshold of 110 to 120 and a 'Fair' threshold of 80. Accepting an 'Excellent' threshold at 110 would suggest a 'Good' threshold approximately halfway at 95 — thresholds at 110, 95, and 80. Accepting an 'Excellent' threshold at 120 would suggest a 'Good' threshold approximately halfway at 100 — thresholds at 120, 100, and 80. None of the 249 river segments in RP River Class 5 are described by the chosen land use filters, suggesting only 'best available' reference sites would be available to validate model predictions and inform 'Excellent' thresholds.

Table 8. Summary Macroinvertebrate Community Index (MCI) statistics for RP River Class 5. *from the contemporary model for sites with native vegetation > 85%, light pasture < 5%, heavy pasture = 0% , urban = 0% , surface water allocation = 0

3.2.9. RP River Class 6

Summary model statistics for RP River Class 6 (Low gradient, small) are provided in Table 9. Data suggest an 'Excellent' threshold of 120 and a 'Fair' threshold of 70. Accepting an 'Excellent' threshold at 120 would suggest a 'Good' threshold approximately halfway at 95 – thresholds at 120, 95, and 70. However, unless these are true soft-bottomed streams (which may be hard to determine) it would be more protective to adopt the proposed national C/D threshold for hard-bottomed streams of 80 (NZFSS submission 4 February 2014). Thresholds would then be 120, 100, and 80.

Table 9. Summary Macroinvertebrate Community Index (MCI) statistics for RP River Class 2. *from the contemporary model for sites with native vegetation > 85%, light pasture < 5%, heavy pasture = 0% , urban = 0% , surface water allocation = 0

3.1. Future work

To validate the reference predictions from current models and to develop predictive models with less error requires MCI data from more sites. Improvements are likely to be achieved by acquiring data from sites currently underrepresented in the training data, such as in GWRC FENZ classes B, C1, C10, C8, UR.

To determine the most appropriate MCI statistic would require an analysis of temporal variability in MCI data from representative reference sites.

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5. APPENDICES

Appendix 1. Summaries of measured and modelled Macroinvertebrate Community Index (MCI) values for the Wellington region.

A1.1. Measured contemporary Macroinvertebrate Community Index (MCI).

A1.2. Boosted regression tree (BRT) modelled contemporary Macroinvertebrate Community Index (MCI).

A1.3. Boosted regression tree (BRT) modelled contemporary Macroinvertebrate Community Index (MCI) at sites with restricted land-use pressure (Natveg >85%, SWA = 0, Urban = 0, HeavyPastoral = 0, LightPastoral <5%)

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A1.4. Linear modelled contemporary Macroinvertebrate Community Index (MCI).

A1.5. Linear modelled contemporary Macroinvertebrate Community Index (MCI) at sites with restricted land-use pressure (Natveg >85%, SWA = 0, Urban = 0, HeavyPastoral = 0, LightPastoral <5%)

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A1.6. Measured reference Macroinvertebrate Community Index (MCI).

A1.7. Boosted regression tree (BRT) modelled reference Macroinvertebrate Community Index (MCI.

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RP $\overline{}$ **GW River Description FENZ Class Class**

A1.8. Linear modelled contemporary Macroinvertebrate Community Index (MCI).

Appendix 2. Boxplots of measured and modelled Macroinvertebrate Community Index (MCI) values for the Wellington region grouped by Greater Wellington Regional Council (GWRC) FENZ class. Note: FENZ = Freshwater Ecosystems of New Zealand geo-database.

A2.1. Boxplots showing measured Macroinvertebrate Community Index (MCI) and contemporary MCI from linear and boosted regression tree (BRT) models predictions grouped by GWRC FENZ Class.

A2.2. Boxplots showing measured Macroinvertebrate Community Index (MCI) and reference MCI from linear and boosted regression tree models predictions grouped by GWRC FENZ Class.